

Toxicity of Methylmercury to Daphnia pulex

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Recognition of the seriousness of polluting intensified when with mercury it was environment established that under anaerobic conditions mercury ions can be methylated to form methylmercury (Wood 1985), a highly toxic and significantly more mobile form of Methylmercury is bioconcentrated in aquatic mercury. organisms in the food chain by direct uptake from the water and biomagnified by ingestion of contaminated foods.

While toxicity of methylmercury to humans and mammals has received much attention, toxicity to aquatic organisms has not been sufficiently examined (D'ltri 1972; Miller & Clarkson 1971; Hartung and Dinman 1972).

Toxic effects of methylmercury on <u>Daphnia pulex</u> were determined, including estimates of acute mortality, and chronic net reproduction (R_{\circ}) survivorship (1) and growth rate. Estimates of LC₅₀ and EC₅₀, as well as accumulated effects over three generations were established.

Thus we have provided a comparison of chronic toxic effects of mercury derived from using the lifetable approach (Seber 1973) with simple effects based on intrinsic growth rates. We have also shown that substantial reductions occur in $EC_{\rm x}$ from one generation to the next.

MATERIALS AND METHODS

Acute tests (LC₅₀) employed <u>D. pulex</u> [The EPA clone is a cross <u>D. pulex</u> x <u>D. pulicaria</u> (P. Hebert pers. comm.] neonates (<24 hr old). Test water was filtered with Whatman qual. 1 filter paper. Ten animals were placed in

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100 mL test water at 20° C. Treatment involved five concentrations of methylmercury (chloride) prepared without solvents, with five replicates each.

Lifetable experiments utilized <u>D. pulex</u> neonates. Test water was again filtered lake water. Experiments ran for 30 days. Four concentrations of methylmercury (chloride) (10ng/L, 100 ng/L, 500 ng/L and 1000 ng/L) were used in experimental treatments. Neonates were placed in 40 mL test water (50 mL beakers) with 20 replicates at 20° C. A photoperiod of 16 hr light/8 hr dark was used. Yeastalfalfa fish-chow suspension was added for food. Progeny were counted and removed daily. Test water was changed every three days.

For growth rate experiments the initial size (L_o) was taken by average size of 20 animals (<24 hr old) randomly chosen from about 200. Treatment concentrations of methylmercury were 0 ng/L (control), 100 ng/L, 500 ng/L, 1000 ng/L and 1500 ng/L. Twenty test animals were raised in filtered lake water(100 mL) with added food suspension as a food resource and the test water was changed every two days. Final size (L_2) was the average size of 20 animals after seven days growth in test water. Size measurements were made from tip of the helmet to base of the tail spine.

The LC_{50} 's and EC_{25} 's and EC_{50} 's (as shown in Table 3) were determined by linear regression (Statistix 3.5 c). Treatment effects on R_o (Table 2) were determined using an unbalanced ANOVA (Statistix 3.5).

Concentrations of methylmercury for acute tests were 1000-45000 ng/L and for chronic tests were 10-1000 ng/L, which were higher than the concentration of natural water [general environmental levels of methylmercury are 8 ng/L (Wood 1985)] but very much lower than those concentrations which cause mercury poisoning in humans [0.2-1.0 ppm (Grant 1971)].

The test animals (\underline{D} . \underline{pulex}) were taken from an EPA clone; test water was collected from Lake Calhoun, Minneapolis. These test conditions were closer to nature than reference water made by addition of artificial ions to distilled water.

Males occurred while using <u>Daphnia</u> lifetables for toxicity tests (Table 1). This biological phenomenon is normal. Young males should be removed at the onset to avoid discarding experiments.

Table 1. The number of males occurring in test groups (N=20)

| Methylmercury concentration | | | | |
|-----------------------------|-------------------------|-------------|---------------------------|--|
| (ng/L) | \mathtt{f}_{\diamond} | ${\bf f}_1$ | $\mathtt{f}_{\mathtt{2}}$ | |
| 0 | 0 | 0 | 2 | |
| 10 | 0 | 0 | 0 | |
| 100 | 0 | 1 | 2 | |
| 500 | 0 | 2 | 1 | |
| 1000 | 0 | 2 | - | |

RESULTS AND DISCUSSION

The concentration of methylmercury affecting survival of <u>Daphnia magna</u> was higher (260-870 ng/L) (Biesinger et al. 1982) than that for <u>D. pulex</u> reported here. Likewise, the EC_{50} for reduction in reproductive rate was higher (1140 ng/L).

The LC₅₀s were 31205 ng/L (24 hr), 5700 ng/L (48 hr) and 1805 ng/L (96 hr). The lethal concentrations of methylmercury illustrate the reduced sensitivity of acute tests.

Experimental results from lifetables are summarized in Table 2 (a,b,c). Treatments from 10 ng/L to 1000 ng/L methylmercury caused gradual reduction in age-specific birth rates (m_x) and net reproductive rate (R_o). Much more pronounced effects were observed at concentrations greater than 100 ng/L.

With regard to the effects of methylmercury on net reproduction, an obvious break between 100 and 1000 ng/L was significant (p<0.95 - p<0.001) (Figure 1). The critical concentration for methylmercury was between 10-100 ng/L (Table 2c). The effective concentration (EC₅₀) calculated from R_o was 707 ng/L (f_o), 336 ng/L (f₂), 229 ng/L (f₂) (Table 3). Effective concentration (EC₅₀) declined from generation 1(f_o) to generation 3 (f₂). The ratio R_o (treatment)/R_o (control) shows clearly the effects of various concentrations of methylmercury by generation (Figure 2).

Some animals died before the age of reproduction (8 days) in all treatments (Table 4); however, the percentage dead gradually increased with increasing methylmercury concentration. The LC_{25} values (8 days) were 794 $ng/L(f_0)$, 275 $ng/L(f_1)$, and 331 $ng/L(f_2)$ (Table 5).

Table 2. Lifetable, providing estimates for natality (m_x) and estimates of net reproduction $(R_o = 1_x m_x)$ of <u>Daphnia pulex</u> exposed to methylmercury for 30 days.

| Concentr | ation | | | | | |
|----------------------|----------|----------------|----------------|------|----------------|------------|
| (ng/L) | <u> </u> | 1 _x | m _× | +SD | R _o | <u>+sD</u> |
| a) f _e ge | nerati | .on | | | | |
| 0 | 20 | 1.0 | 119.5 | 7.4 | 119.5 | 16.1 |
| 10 | 20 | 1.0 | 116.9 | 7.1 | 116.9 | 10.3 |
| 100 | 20 | 1.0 | 110.8 | 6.3 | 110.7 | 13.5 |
| 500 | 20 | 0.94 | **91.9 | 6.8 | *86.5 | 12.9 |
| 1000 | 20 | 0.64 | ***63.9 | 10.6 | ***40.6 | 5.7 |
| | *** | | | | | |
| b) f_1 | genera | tion | | | | |
| 0 | 20 | 0.97 | 98.5 | 9.3 | 98.4 | 11.9 |
| 10 | 20 | 0.97 | 94.5 | 9.9 | 92.4 | 14.9 |
| 100 | 19 | 0.93 | 85.1 | 12.6 | 79.3 | 15.6 |
| 500 | 18 | 0.66 | ***58.6 | 4.9 | 38.7 | 8.4 |
| 1000 | 18 | 0.58 | ***43.9 | 8.1 | ***25.5 | 6.0 |
| | | | | | | |
| c) f_2 | genera | tion | | | | |
| 0 | 18 | 0.95 | 134.6 | 5.1 | 127.4 | 16.8 |
| 10 | 20 | 0.95 | 111.6 | 10.2 | 106.0 | 14.7 |
| 100 | 18 | 0.85 | 115.0 | 5.7 | *98.7 | 13.2 |
| 500 | 19 | 0.80 | ***33.7 | 8.1 | ***25.5 | 6.0 |
| | | | | | | |
| * 0.0 | 0>q>1(| . 05 | | | | |

^{* 0.01&}lt;p<0.05

p<0.001

Both the LC_{50} for methylmercury and the EC_{50} for reduction in reproductive rate were lower (87-114 ng/L) than previously reported.

Table 3. EC_{25} and EC_{50} derived from R_{\circ} of <u>Daphnia pulex</u> exposed to methylmercury.

| | f。 | f ₁ | f_2 | mean | |
|------------------|-----|----------------|-------|------|--|
| EC ₂₅ | 380 | 131 | 110 | 207 | |
| EC ₅₀ | 707 | 336 | 229 | 424 | |

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p<0.01

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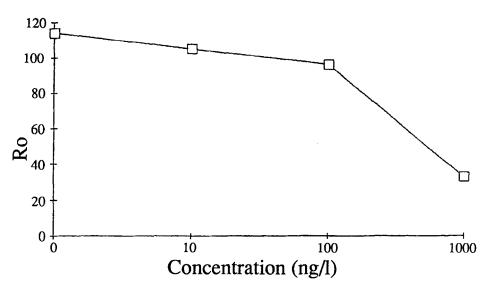


Figure 1. The relationship between $R_{\rm o}$ and concentration of methylmercury

Table 4. Percent death of <u>Daphnia pulex</u> before age of reproduction (8 days, when exposed to methylmercury (ng/L).

| Methylmercury concentration | f _o | f ₁ | f ₂ | |
|-----------------------------|----------------|----------------|----------------|--|
| 0 | 0 | 0 | 5 | |
| 10 | 0 | 0 | 5 | |
| 100 | 0 | 15 | 10 | |
| 500 | 5 | 30 | 30 | |
| 1000 | 35 | 35 | _ | |

Table 5. Values of LC_{25} for reproduction (8 days) with concentration of methylmercury (ng/L).

| | f。 | f_1 | f ₂ | mean |
|------------------|-----|-------|----------------|------|
| LC ₂₅ | 794 | 275 | 331 | 467 |

<u>Daphnia</u> intrinsic growth rates (Table 6a) declined gradually with increasing methylmercury concentration. EC_{25} of 1258 ng/L was much higher than the critical concentration observed in lifetable experiments (207 ng/L). The sensitivity of growth tests (r) is therefore less than lifetable tests (R_{\circ}).

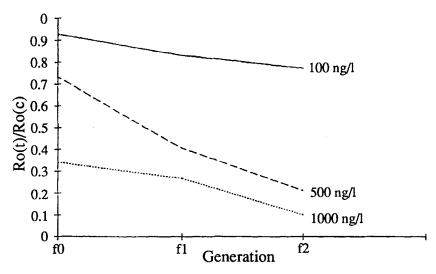


Figure 2. Ratios of R_o (treatment)/ R_o (control) for each generation.

Table 6a. Intrinsic growth rate [r=ln $(l_1/L_o/t_2-t_o)$], as influenced by methylmercury concentration. [N=number of animals, (t_2-t_o) =7 days].

| Treatment concenta (ng/L) | ation | chylmercury r (replicate ₁) | r (replicate ₂) | r (replicate ₃) |
|---------------------------|-------|---|--------------------------------|--------------------------------|
| 0 | 20 | 0.106 | 0.098 | 0.082 |
| 100 | 20 | 0.109 | 0.098 | 0.082 |
| 500 | 20 | 0.091 | 0.087 | 0.079 |
| 1000 | 20 | 0.084 | 0.082 | 0.072 |
| 1500 | 20 | 0.070 | 0.060 | 0.058 |

L_o = starting size

 $L_1 = final size$

Table 6b. The value of EC₂₅ (derived from growth rate) with concentrations methylmercury (ng/L).

| Groups | 1 | 2 | 3 | mean |
|------------------|------|------|------|------|
| EC ₂₅ | 1148 | 1310 | 1318 | 1258 |
| | | | | |

Using chronic estimates of mortality and reproduction, we have established that chronic effects occurred at lower concentrations (580-794 ng/L) than acute effects [LC $_{50}$ =5700 ng/L (48h)]. The acute/chronic ratio was 9.7:1. The US-EPA has suggested (R. Carlson, pers. Comm.) a probable ratio of 25:1. The determination of LC $_{\rm x}$ and EC $_{\rm x}$ with methylmercury on <u>D. pulex</u> will provide

a bioassay with which to estimate the concentration of methylmercury in natural waters over the world. Such techniques will be eagerly applied in China.

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